

Assessment of Metal - Induced Inhibition of Soybean Urease as a Tool for Measuring Heavy Metals in Aqueous Samples

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Abstract : Inhibition of soybean (*Glycine max*) urease by heavy metal ions has been investigated with a view to developing a method for the indirect determination the heavy metals. Urease activity was assayed as a function of metal concentration in the presence of Cd²⁺, Cr³⁺, As³⁺, Cu²⁺, Pb²⁺ and Zn²⁺. Our results revealed concentration dependent inhibition of urease activity within the range of 0.001 – 10 mg/L for Cu²⁺, As³⁺, Cr³⁺ and Cd²⁺ and 0.1 – 10 mg/L for Zn²⁺ and Pb²⁺. The inhibitory strengths of the metals as evaluated from their IC50 values showed the following ranking: Cu²⁺ > As³⁺ > Cr³⁺ > Cd²⁺ > Zn²⁺ > Pb²⁺. Assays for multi-metal solutions revealed similar findings as the single-metal systems but with higher inhibitory strengths. The calibration plots of % inhibition against concentration displayed different linear ranges for the investigated metals; Cu As, Cr and Cd showed linearity within the range of 0.01 – 10mg/L while Zn and Pb plots were linear within 0.1 – 10 mg/L. The applicability of the assay for the quantitative determination of heavy metals was evaluated by analysing synthetic water samples and comparing the results with a standard method (AAS). T-test examination revealed 74% agreement between the two methods. The results of this study show that soybean urease can be valuable for the inhibitive determination of heavy metals in aqueous samples.

Keywords: Enzyme, Heavy metals, Inhibition soybean, urease

I. Introduction

In recent years, the use of enzymes for analytical purposes has gained considerable importance in the quest for simple, rapid and inexpensive methods for real-time determination of compounds that may adversely affect human health and the environment. Enzymes as catalysts make excellent analytical reagents due to their selectivity and specificity and are frequently used to determine the concentration of their substrates. On the other hand, inhibition of specific enzymatic systems is also being applied as analytical methods for the detection of toxic pollutants such as pesticides [1, 2] herbicides [3] and heavy metals [4, 5]. Heavy metal ions constitute a serious environmental problem due to their ubiquitous and non-biodegradable nature. The metals and their compounds are notable for their abundance, persistence, high toxicity and ability to accumulate in living organisms [6]. The toxic effect of heavy metals on biological systems stems from their interference with the normal metabolic processes of bio-molecules. The metals tend to replace other metal groups or hydrogen atoms in bio-molecules such as proteins and enzymes to form strong and stable chemical bonds. These interactions affect the normal functioning of the molecules and in the case of enzymes can lead to inhibition or complete inactivation of catalytic properties. The inhibition of enzymes by heavy metals has attracted great attention in the area of bio-monitoring of the pollutants and many enzyme- based systems have been reported for the inhibitory analysis of heavy metal ions.

Owing to its pronounced sensitivity, urease (urea amidohydrolase, EC 3.5.1.5), has been extensively applied as a probe for heavy metal ions. The enzyme is responsible for the hydrolysis of urea into carbon dioxide and ammonia [7]. It is abundantly present in different seeds, microorganisms as well as in certain invertebrates [8]. The structure and molecular weight of urease depend on the enzyme origin. In plants, it is a hexamer that consists of six identical chains and is located in the cytoplasm [9]. In the case of soybeans, the urease enzyme plays an important role during the process of germination where the protein stored within the seed is mobilized in order to nourish the seedlings [10]. Two urease isoenzymes; a tissue-ubiquitous and embryo-specific have been identified in soybean [11, 12]. The embryo-specific urease is an abundant seed protein while the other type (called ubiquitous) is found in lower amounts in vegetative tissues. The sensitivity of urease to heavy metal ions is due to the presence of multiple cysteine residue of which one, conserved principally to all known ureases, is located in the mobile flap of the enzyme active site [13]. Heavy metal ions inhibit the catalytic activity of urease by binding with a sulfhydryl group in the active center of the enzyme in a reaction analogous to the formation metal sulphides [14]. Once the metal atom is bound, the sulfhydryl group cannot function in catalysis hence there will be a marked decrease in the catalytic activity. Since very low ratios of inhibitor compounds to enzyme molecules can have a dramatic effect on the enzymatic activity, the indirect determination of trace concentrations of the inhibitor is possible by monitoring the enzymatic activity. This study is aimed at investigating the inhibitory effects of selected heavy metals on *G. max* urease with a view to

evaluating its usefulness for the inhibitive determination of the metal ions. Although Soybean seeds are known to be an important source of urease, there a dearth of information on its use in enzyme - based analytical systems. This study will uncover the potential of soybean urease and provide data that will allow further exploitation in the area of heavy metal analysis.

II. MATERIALS AND METHODS

2.1 Materials

All chemicals were prepared using distilled water. The buffers used for enzyme inhibition tests were Sodium hydrogen phosphate, sodium dihydrogen phosphate obtained from BDH chemicals. Ammonium sulphate, Urea and sodium hydroxide were obtained from JDH chemicals. KNO₃ was from Burgoyne, India and HCl from Merck, Germany. Nessler's reagent was prepared from potassium iodide (Qualikems, Nigeria) and mercuric chloride (Sigma-Aldrich). The standards for cadmium (II), Arsenic (III) and Chromium (III) were purchased as atomic absorption standard solutions (1000 mg/L) (Sigma-Aldrich). Analar grades of the salts Cu(NO₃)₂.3H₂O, Pb(NO₃)₂, Zn(NO₃)₂.6H₂O from BDH chemicals were used to prepare stock solutions of the metals. Soybean seeds were purchased from the local market in wukari, Nigeria and the urease extracts were prepared freshly each day.

2.2 Extraction of urease

Powdered soybean seeds (10 g) were soaked overnight in 100 mL extraction buffer (0.2 M sodium phosphate buffer pH 7.3) at 4 °C (Refrigeration temperature). The mixture was sieved through four layers of muslin cloth and the resulting filtrate was centrifuged at 4000 rpm for 30 minutes. The clear supernatant was collected and used as the crude urease extract (stored at 4 °C) while the pellets were discarded.

2.3 Determination of urease activity

The activity of the extracted crude urease was estimated using a slightly modified method in which ammonia released is spectrometrically determined [15]. To 0.8 mL of assay buffer (0.2 M phosphate buffer, pH 7.3), 0.2 mL of enzyme extract was added and incubated at 30°C with 1 mL of urea solution (0.25M in phosphate buffer) . After 10 minutes, the reaction was terminated by addition of 1 ml of 0.1 M HCl. The reaction mixture was made up to 50 mL with distilled water and 2 mL of Nessler's reagent was added. Absorbance of the resulting solution was read against a blank at 405 nm on a UV-Vis spectrophotometer (APEL PD 3000). Urease activity was estimated (as μM of ammonia released) from an ammonium sulphate standard curve. One enzyme unit is defined as the amount of enzyme required to liberate 1 μM of ammonia per minute under the test conditions (0.25M urea, 0.1M buffer, pH 7.3 at 30°C).

Urease activity of the sample was calculated using equation (1)

$$\text{Urease Activity (U/mL)} = \frac{A(\text{sample}) \times \text{dilution factor}}{\text{Slope} \times T \times v} \text{----- (1)}$$

Where: A (sample) = measured absorbance of the Sample against the blank

T = incubation time (10 minutes for standard assay).

V = volume of sample

2.4 Determination of Kinetic parameters (Km and Vmax) of the extracted urease

The kinetics of the extracted urease was evaluated using the Michealis - menten model. The activity of the enzyme was studied by varying the amount of substrate (urea) from 0 to 40 mM at constant enzyme concentration. The V_{max} (i.e., maximum velocity achieved by the system, at saturating substrate concentrations and Michaelis-Menten constant, K_m (the substrate concentration at which the reaction velocity is 50% of the V_{max}) were estimated from the Michelis-Menten plot of reaction velocity vs substrate concentration.

2.5 Inhibition Studies

Standard solutions of the metal ions; Cu²⁺, Cd²⁺, Zn²⁺, Cr³⁺, As³⁺ and Pb²⁺ and multi – metal solutions with concentrations ranging from 0.001 – 100 mg/l were prepared by serial dilution from their respective 1000 ppm stock and their inhibitory activities were then tested on the *G. max* urease. In the procedure adopted for this study, 0.2 ml metal ion solution, 0.2 ml enzyme extract and 0.6 ml buffer (pH 7.3) was allowed to incubate at 30° C for ten minutes to allow adequate time for metal/enzyme interaction. 1mL of urea (0.25M) was then added and after another 10 minutes, the reaction was terminated by adding 1 ml of 0.1 M HCl. The reaction mixture was made up to 50 ml with distilled water and 2 mL Nessler's reagent was added. Absorbance of the resulting solution was read against a blank at 405 nm on a UV-Vis spectrophotometer (APEL PD 3000). The level of inhibition for each tested metal concentration was obtained using equation (2).

$$\% \text{ inhibition} = [A_0 - A_I] / A_0 \times 100 \text{ ----- (2)}$$

Where; A_0 = the obtained absorbance without the inhibitor (metal ion)

A_i = absorbance obtained after pre- incubation with metal ion.

A calibration graph of % inhibition against concentration was plotted for the single and multi-metal solutions

2.6 Determination of Heavy Metals in Synthetic water Samples Using Urease Inhibition Assay

Ten (10) synthetic water samples were prepared for each of the metals: Cu, As, Zn, Cr, Pb and Cd by adding unknown quantities of the metal solutions to distilled water. Multi-metal synthetic samples were similarly prepared by adding unknown quantities of the various metals to distilled water. All the prepared synthetic samples were subjected to inhibition assays as described in 2.5. The concentrations of the various metals in the samples were estimated from the regression equations of the respective inhibition calibration curves.

2.7 Determination of Heavy metals in Synthetic Samples Using Atomic Absorption Spectroscopy

Heavy metal analysis of the synthetic water samples was done with atomic absorption spectrophotometer AA-6800 (Shimadzu, Japan). A calibration curve of absorbance against concentration was prepared for each metal by running different standard solutions of the metals and the concentrations of the samples were estimated by extrapolation from the respective calibration curves.

2.8 Comparison of Inhibition Assay Method with Atomic Absorption Spectroscopy

The results of the synthetic water analysis obtained via the urease inhibition assay and AAS were compared through 't' test analysis. [Components: metal type (Cr, Cu, Cd, As, Zn, Pb and mixed metals); Methods (AAS, and spectrophotometric inhibition assay).]

III. Results and Discussions

3.1 Enzyme activity and kinetic Characteristics

The urease used in this study which was extracted from soybean (*G.max*) seeds revealed an enzyme activity of 15.065 IU/mL. One IU (international unit) of urease activity is the amount of urease that produces 1 μ mole of ammonia from urea in one minute under the stated test conditions. The enzyme kinetics as evaluated via the Michaelis – Menten model showed a simple Michaelis-Menten type kinetic behaviour in accordance with what has been described for ureases from other sources. The K_m value obtained by hyperbolic regression yielded 5.95 ± 1.03 mM, and V_{max} was found to be 0.45mM/min. The K_m is a measure of how well a substrate complex with a given enzyme, otherwise known as its binding affinity. An equation with Low K_m value indicates a large binding affinity therefore the reaction will approach V_{max} more rapidly. A high K_m indicates that the enzyme does not bind efficiently with the substrate, and V_{max} will only be reached if the substrate concentration is high enough to saturate the enzyme. *G. max* urease in this study shows a higher affinity for its substrate than the ureases from *B. abortus*, 13 mM [16], *B. badius*, 7.69mM [17] and *Canavlia ensiformis*, 19.10 mM [18] but less affinity than jack beans urease, 4.6mM [19], *Phaseolus vulgaris* (2.1 mM) [20] and *H. pylori* whose urease has one of the lowest K_m values (0.3 mM) [21]. The findings here, suggest that the extracted urease has good affinity for the substrate (urea) under the conditions tested.

3.2 Inhibition Studies on Single – metal solutions

Six heavy metal ions; Cu^{2+} , As^{3+} , Zn^{2+} , Pb^{2+} , Cd^{2+} and Cr^{3+} were tested to evaluate their effects on *G.max* urease activity in aqueous environments. Typically the interaction of heavy metals with urease leads to decreased catalytic activity due to the binding of the metal ions with the enzyme thiol groups [14]. The results of the inhibition assays (Tables 1 – 6) show that the metals examined in this study had different effects on urease activity. The obtained absorbances reflect the quantity of ammonia released from the catalytic activity of urease (the hydrolysis of urea to HCO_3^- and NH_4^+). The control represents the enzyme activity in the absence of heavy metals. The assay for Cu (II) ions (Table 1) revealed decreased responses with increased concentration corresponding to increased inhibition with increasing concentration. As can be seen, there is a dose dependent inhibition of urease activity by Cu^{2+} within the range 0.01 – 10 mg/L. The data on Arsenic (As^{3+}) assay also established an inhibition range of 0.001 – 10 mg/L. 24.27% inhibition was observed at 0.001mg/L concentration while 88.59% was recorded at 10 mg/L. The inhibitory effect of Arsenic ions on *G. max* urease activity was stronger than that of Cu ions. This can be clearly seen from the higher percentage inhibition exhibited by arsenic at all the tested concentrations. This strong inhibition by As^{3+} could be due to its toxic nature. Arsenic is a highly toxic metalloid which ranks first on the Environmental Protection Agency (EPA) Priority List of Hazardous Substances [22]. Inhibition of urease by heavy metal ions has been related to its biological toxicity, which causes loss of the biological function of the enzyme [23]. Most studies published on inhibition of urease

of both plant and bacterial origin by heavy metal ions have been aimed either at investigating their toxicity or at estimating their concentrations [24]. Arsenic ions have been shown to inactivate enzymes by reacting with the enzyme sulphhydryl groups which results in the formation of arsenic sulphide [13]. It has also been reported that As^{3+} decreases enzyme activity in three ways: (1) by interacting with the enzyme–substrate complex; (2) by denaturing the enzyme protein or; (3) interacting with the active protein groups [25].

Table 1: Urease Inhibition Assay for Copper

Cu^{2+} Concentration (mg/L)	Absorbance (at 405 nm)	Inhibition (%)
0.00 (control)	1.606 ± 0.03	0
0.001	1.416 ± 0.02	11.83
0.01	1.313 ± 0.11	18.24
0.1	1.252 ± 0.04	22.04
1	1.137 ± 0.12	29.20
10	0.202 ± 0.05	87.42

Absorbances are presented as mean ± SD for three determinations

Table 2: Urease Inhibition Assay for Arsenic

As^{3+} Concentration (mg/L)	Absorbance at 405 nm	Inhibition (%)
0.00 (Control)	1.718 ± 0.04	0
0.001	1.301 ± 0.09	24.27
0.01	1.007 ± 0.12	41.38
0.1	0.967 ± 0.06	43.71
1	0.889 ± 0.11	48.25
10	0.196 ± 0.03	88.59

Absorbances are presented as mean ± SD for three determinations

The inhibition assay for zinc revealed some interesting findings. At the lowest tested concentration of 0.001mg/L, contrary to expectations, Zn ions caused an increase in urease activity. However, at the next tested concentration (0.01mg/L) Zn^{2+} did not result in higher activation of the enzyme. The inhibition tests revealed an absorbance of 1.716 at 0.001mg/L and 1.681mg/L at 0.01 mg/L while the control had an absorbance of 1.652. These values indicate higher enzyme activation at a lower concentration (0.001 mg/L) than at 0.01 mg/L. The decreased activation at 0.01 mg/L suggests that partial activation and inhibition occurred concurrently at that concentration. Low concentrations of certain metal ions have been shown to enhance urease activity; Turgay and Namli, [26] reported increased urease activity at low concentrations of Cd and Pb. Similar trend has also been observed with Nickel ions [27]. Our findings here demonstrate that low concentration of Zn (II) ions can enhance the catalytic activity of *G.max* urease. The presence of low concentrations of heavy metal ions produces stress promoters in plant cells which can lead to increased enzyme activity. As long as the stress is not too strong for the plant’s defence capacity, the response to the heavy metals is an increase in the activities of certain enzymes [28]. Also, in some cases, heavy metals at low concentrations may act as cofactors and activators to enhance the enzyme activity [29]. Inhibition of urease activity by Zn^{2+} was observed within the range of 0.1 – 10 mg/L. Compared to Cu^{2+} and As^{3+} , Zn^{2+} demonstrated weaker inhibitory effects on urease activity. Zn ion has been previously reported as a relatively weak inhibitor of urease [30]. This is probably due to its less toxic nature. Treatment of *G.max* urease with Chromium (III) ions also led to a dose dependent inhibition of the enzyme activity. The inhibition was observed over a concentration range of 0.001-10 mg/L. The assay revealed an inhibition of 8.84% at 0.001 mg/L and 78.45% at 10 mg/L concentration. Cr^{3+} showed stronger inhibitory effects than Zn^{2+} but weaker than Cu^{2+} and As^{3+} ions. These trends can be justified either in terms of their respective affinities for urease -SH groups or their relative toxicities.

Table 3: Urease Inhibition assay for zinc

Zn^{2+} Concentration (mg/L)	Absorbance (at 405 nm)	Inhibition (%)
0.00 (Control)	1.652 ± 0.06	0
0.001	1.716 ± 0.12	NI
0.01	1.681 ± 0.13	NI
0.1	1.406 ± 0.07	14.89
1	1.311 ± 0.11	20.64
10	0.602 ± 0.03	63.55

Absorbances are presented as mean ± SD for three determinations; NI=No inhibition

Table 4: Urease inhibition assay for chromium

Cr ³⁺ Concentration (mg/L)	Absorbance (at 405 nm)	Inhibition (%)
0.00 (Control)	1.652 ± 0.03	0
0.001	1.506 ± 0.07	8.84
0.01	1.457 ± 0.10	11.08
0.1	1.406 ± 0.05	14.89
1	1.256 ± 0.11	23.97
10	0.356 ± 0.13	78.45

Absorbances are presented as mean ± SD for three determinations

Data on the inhibition assay for Cd²⁺ (Table 5) show that the catalytic activity of *Glycine max* urease is inhibited in a dose- dependent fashion by Cd ions within the range of 0.001-10 mg/L concentration. The inhibition caused by cadmium appear to be less than that observed for Cu²⁺, As³⁺ and Cr³⁺ but greater than the inhibition by Zn²⁺. This suggests that perhaps Cd²⁺ has a lower affinity for urease thiol groups than the three metals (Cu, As and Cr). Lead (II) ions showed weak inhibitory effects on *G.max* urease activity as seen from the results of the inhibition assay presented in Table 6. No clear inhibition of enzyme activity was observed at 0.001 and 0.01 mg/L Pb²⁺ concentration. However at 0.1 mg/L, a pronounced decrease in the activity was recorded. Percentage inhibition range of 13.07 - 57.93 was obtained over a concentration range of 0.1-10 mg/L. The inhibition caused by Pb was much less than observed for the other metals investigated. A Survey of published data confirms the weak inhibitory effects of Pb²⁺ on urease activity; It has been reported that Pb (II) ions did not produce a significant decrease in urease activity at concentrations up to 50 mg L⁻¹ [31], whereas Cd (II) and Zn(II) inactivated urease in the range of 1–30 mg L⁻¹. [32]; Lee and Lee [33] also showed that Pb²⁺ exhibited weaker inhibition of urease activity compared with Cu²⁺ and Cd²⁺.

Table 5: Urease inhibition assay for cadmium

Cd ²⁺ Concentration (mg/L)	Absorbance at (405 nm)	Inhibition (%)
0.00 (Control)	1.658 ± 0.06	0
0.001	1.475 ± 0.01	11.03
0.01	1.339 ± 0.10	19.24
0.1	1.288 ± 0.05	22.32
1	1.198 ± 0.01	27.74
10	0.421 ± 0.03	74.61

Absorbances are presented as mean ± SD for three determinations

Table 6: Urease inhibition assay for Lead

Pb ²⁺ Concentration (mg/L)	Absorbance (at 405 nm)	Inhibition (%)
0.00 (Control)	1.652 ± 0.06	0
0.001	1.648 ± 0.09	0.24
0.01	1.650 ± 0.12	0.12
0.1	1.436 ± 0.06	13.07
1	1.375 ± 0.10	16.77
100	0.695 ± 0.07	57.93

Absorbances are presented as mean ± SD for three determinations

3.3 Inhibition Assay for mixed - metal Solutions

Co-existence of heavy metals is commonly encountered in the environment as a result of human and natural activities. During inhibition analysis, the presence of other metal ions may elicit antagonistic, additive or synergistic effects. In order to understand the effect of these chemicals in totality, and also ascertain the applicability of the proposed system for sensing of total heavy metal ions, assays were carried to evaluate the inhibition of *G.max* urease by mixtures of all the investigated heavy metals. The results of the inhibition assays on the mixed- metals (Table 7) also demonstrated dose dependent inhibition of urease activity. However, the measured responses were found to be greater than for the single-metal solutions at all the tested concentrations. The inhibition pattern seemed synergetic but not additive. The effects of a mixture of metal ions on urease activity have been previously shown to be synergetic in such a way that the total urease inhibition can be calculated by addition of the estimated inhibition values for individual cations [34]. Our finding here does not indicate such additive synergism but suggests a potentiating type. Potentiation here refers to the enhancement of one agent by another so that the combined effect is greater than the effect of each one alone. The inhibition pattern demonstrated by the mixed-metal solutions reflects such effects. For the single metal systems, the inhibition obtained at the lowest tested concentration (0.001 mg/L) ranged from 8.84 -24.27% while for the mixed - metal systems, 37.97% was obtained. At 10 mg/L, 55.41- 88.59 % inhibitions were obtained for the

single-metal solutions and 97.00% for the multi-metal solutions. These values show that the mixed -metals exhibit a pattern of inhibition enhancement that is not of an additive nature.

Table 7: Urease inhibition assay for mixed metal standards

Concentration (mg/L)	Absorbance (at 405nm)	Inhibition (%)
0.00 (Control)	1.667 ± 0.07	0
0.001	1.034 ± 0.13	37.97
0.01	0.984 ± 0.12	40.97
0.1	0.917 ± 0.16	44.99
1	0.850 ± 0.10	49.01
10	0.05 ± 0.03	97

Absorbances are presented as mean ± SD for three determinations

3.4 Inhibition Plots

In enzyme inhibition assays, a plot of % inhibition against concentration of inhibitor is usually employed as calibration graph for determination of the inhibitor concentration. The features of the inhibition calibration plots for the investigated metal solutions are presented in Table 8. The inhibition plots for Cu²⁺ and As³⁺ showed linear relationships within the concentration range of 0.01 – 10 mg/L. However, the inhibition pattern of arsenic (III) ion was quite different from that exhibited by Copper. Within the same concentration range of 0.001 – 10 mg/L, Cu showed a wider percentage inhibition range; 11.83 – 87.42 against 24.27 – 88.59 obtained for arsenic. These differences are seen reflected in their respective inhibition graphs; Arsenic inhibition plot revealed a smaller slope values (4.575) than copper with 6.6997. The slope is a measure of the sensitivity of a calibration curve; the steeper the slope, the more sensitive the procedure i.e., the stronger the response on y-axis to a concentration change. The lower slope values obtained for arsenic plot indicate lesser steepness which suggests that the system is less sensitive for detection of arsenic than for copper. High sensitivity in inhibition assays is usually due to high inhibition efficiency of the analyte [35]. The inhibition pattern for zinc was linear in the concentration range of 0.1 -1 mg/L with higher slope values than for As³⁺ but lower than obtained for Cu²⁺. This suggests that the system’s sensitivity towards zinc is higher than that of As³⁺ but lower than for Cu²⁺. The inhibition plot for Cr³⁺ exhibited a linear relationship within 0.01-10 mg/L concentration with a correlation of 99.73%.The sensitivity of the inhibition assay for Cr³⁺ can be inferred from the slopes of the calibration lines which was obtained as 6.4241.This value suggest relatively good sensitivity towards Cr; comparable to that of Copper but better than for Arsenic and Zinc. Linearity was established between Cd²⁺ ion concentration and % inhibition of urease activity within 0.01 – 10 mg/L concentration. The slope here also suggests relatively good calibration sensitivity for Cd detection. The linear range for estimation of Pb²⁺ was established between 0.1–10 mg/L concentration with an R² value of 99.99%. It seems that better correlation of the linear model is achieved when the range of inhibition is smaller. Among the investigated metals, Zn²⁺ and Pb²⁺ which showed smaller inhibition range (within 0.1–10 mg/L), consequently gave higher correlation values. The inhibitory responses for the multi- metal solutions were linear in the concentration range of 0.001- 10mg/L. The linear range obtained here is wider than for all the single metal systems. This is a reflection of the intensified inhibitory effects of the mixed metals.

Table 9: Features of inhibition calibration plots for investigated heavy metals

Heavy metals	Regression equation	Linear range	R ² (%)
Cu	Y = 6.6997x + 20.6166	0.01 – 10 mg/L	99.68
As	Y = 4.5975x + 42.7129	0.01 – 10 mg/L	99.79
Zn	Y = 4.8554x + 15.0617	0.1 – 10 mg/L	99.93
Cr	Y = 6.4641x + 14.1437	0.01 – 10 mg/L	99.28
Cd	Y = 5.3677x + 21.0687	0.01 – 10 mg/L	99.72
Pb	Y = 4.5483x + 12.4278	0.1 – 10 mg/L	99.99
Mixed-metal	Y = 5.5569x + 41.6395	0.001 – 10 mg/L	98.94

3.5 Half Maximal Inhibitory Concentration (IC50) Values

The heavy metals investigated in this study showed varying inhibitory strengths on urease activity. This can be seen from their respective IC 50 values (Table 8) derived from least squares linear regression of the % inhibition plots. IC 50 (inhibitory concentration 50) is the concentration of an inhibitor at which 50% inhibition occurs. It is commonly used as a measure of the inhibitory effectiveness of a substance; the lower the IC 50 value, the stronger the inhibitory effect. Our data showed the following ranking of inhibition by the metals: As(III) > Cu(II) > Cr(III) > Cd(II) > Zn(II) > Pb(II). A survey of previously reported studies showed some disagreements on the inhibitory effects of heavy metals on urease activity; The activity of urease has been shown to be inhibited by heavy metals in the order; Cd²⁺ > Ni²⁺> Cu²⁺>Zn²⁺> Co²⁺> Fe³⁺> Pb²⁺ [36], Cr(VI)>Ni(II)>Cu(II)>Cd(II)>Zn(II)>Pb(II) [37] and Hg²⁺> Cu²⁺> Zn²⁺> Cd²⁺> Ni²⁺> Pb²⁺> Co²⁺> Fe³⁺> As³⁺ [30]. The inhibition orders have also been previously reported as Hg²⁺ >Ag⁺ >Cu²⁺>Ni²⁺ > Cd²⁺>Zn²⁺>Co²⁺

>Fe²⁺>Pb²⁺>Mn²⁺ [38] and Cr > Cd > Zn > Mn > Pb [39]. The reasons for the disparities in the reported data could be due to the nature of the ureases used in the investigations. Ureases from different sources are known to have different characteristics and kinetic properties, this factor may be responsible for the differences in inhibitory strengths reported by various studies.

Table 8: Half Maximal Inhibitory Concentration (IC50) Values of the investigated metals

Heavy metals	As ³⁺	Cu ²⁺	Cd ²⁺	Cr ³⁺	Zn ²⁺	Pb ²⁺
IC 50 (mg/L)	1.584	4.386	5.390	5.547	7.197	8.260

3.6 Application of Urease Inhibition Assay to synthetic water samples

An attempt has been made at estimating the heavy metal concentration of some synthetic water samples by means of the urease inhibition assay. The analysis is aimed at evaluating the performance of the assay by comparing with a standard method (atomic absorption spectroscopy). 10 synthetic samples for each metal were prepared by adding unknown quantities of the metal solutions to distilled water. Solutions of mixed metals were also prepared in the same manner. Measurements of urease inhibition by the solutions were carried out and the concentrations of the samples were estimated from their respective inhibition plots. The data from the inhibition assay is summarized in Table 9, the results of the AAS analysis for the multi-metal samples are presented in Table 10 while a comparative evaluation of the results obtained via the two methods is presented in Table 11.

Table 9: Summary of data on urease inhibition assay for heavy metals in synthetic water samples

Heavy metals	Sample size	Absorbance (range)	%inhibition (range)	Concentration range obtained (mg/L)
Copper(II)	10	1.034 – 1.352	15.82 – 35.61	0.278 – 2.238
Arsenic(III)	10	0.805 – 0.858	50.05 – 53.14	1.596 – 2.268
Cadmium(II)	10	1.021 – 1.297	21.77 – 38.41	0.131 – 3.231
Zinc(II)	10	0.486 – 1.268	23.24 – 70.58	1.684 – 11.434
Chromium (III)	10	1.213 – 1.505	8.90 – 26.57	0.106 – 1.804
Lead(II)	10	0.163 – 1.626	3.15 – 90.13	1.007 – 17.084
Mixed metals	10	0.561 – 0.800	52.02 – 66.35	(1.886 – 4.447)*

*Concentrations for samples diluted to a factor of 10⁻¹

The analysis of Cu synthetic samples via the inhibition assay revealed a concentration range of 0.278 – 2.238 mg/L while AAS analysis gave 0.006 – 2.271 mg/L. Statistical comparison of the results by t-test analysis showed good agreement between the two methods; 90% of the results evaluated showed no significant difference from each other at 95% confidence level. Determination of Arsenic by the two methods also gave comparable results; 1.596 – 2.268 mg/L and 1.553 – 2.292 mg/L from the inhibition assay and AAS methods respectively. Only 10% significant difference was observed among the results. The concentration of cadmium obtained from AAS analysis was in the range of 0.140 – 3.216 mg/L against 0.131 – 3.231 mg/L obtained from the inhibition method. T-test evaluation showed that only one out of the 10 samples investigated differed significantly at 95% confidence level.

Table 10: Determination of heavy metals in mixed-metal synthetic samples by Atomic Absorption spectroscopy (AAS)

Sample	Obtained concentration (mg/L)						Total
	Cu ³⁺	As ³⁺	Cd ²⁺	Cr ³⁺	Zn ²⁺	Pb ²⁺	
MM-A	4.613±0.14	0.137±0.05	4.889±0.13	0.454±0.02	19.490±1.2	1.672±0.11	31.185±1.65
MM-B	2.280±0.12	0.134±0.07	7.405±0.22	0.643±0.10	9.469±0.56	0.945±0.09	20.876±1.16
MM-C	4.899±0.25	0.157±0.04	16.64±1.12	0.625±0.08	17.138±1.05	3.419±0.20	42.878±2.74
MM-D	2.167±0.15	0.121±0.02	8.546±0.42	0.365±0.03	18.086±0.62	2.773±0.23	32.058±1.47
MM-E	1.113±0.07	0.180±0.01	24.037±0.72	0.903±0.06	16.410±0.46	2.489±0.05	45.132±1.37
MM-F	5.659±0.05	ND	8.806±0.32	0.471±0.00	17.398±0.24	3.016±0.18	35.353±0.79
MM-G	5.679±0.10	ND	13.618±0.21	0.221±0.01	12.091±0.13	4.237±0.08	35.546±0.52
MM-H	7.732±0.13	ND	16.091±0.51	0.496±0.02	18.854±0.87	4.655±0.03	47.827±1.56
MM-I	4.313±0.07	0.008±0.00	20.018±0.62	0.938±0.12	14.338±0.14	2.003±0.07	41.618±1.02
MM-J	9.383±0.36	ND	7.165±0.16	0.123±0.03	18.062±0.88	2.800±0.12	37.533±0.55

Results are presented as Mean ± SD for three determinations

The data on the estimation of Cr (III) by the inhibition assay method gave 0.118 – 1.910 mg/L while the AAS method gave 0.108 – 1.770 mg/L concentration. Two out of the 10 investigated samples could not be quantified by the two methods implying that the concentrations were below the detection limits of both procedures. Although the inhibition assays established some levels of inhibition for the two samples (Cr-D and Cr-F), they could not be quantified as the obtained inhibition values were not within the linear range of the Cr calibration plot. T-test examination revealed that the concentrations of two samples differed significantly among the two methods. AAS analysis for Zn gave a concentration range of 1.788 – 11.625 mg/L and the inhibition assay gave 1.684 – 11.434 mg/L. The results from the two techniques were not significantly different for 80% of the analysed samples at 95% confidence level. Pb concentration in the range of 0.007 – 17.43 mg/L was

obtained via AAS while 1.007 – 17.084 mg/L was obtained via the urease inhibition assay. Here, the inhibition method differed significantly from AAS in 40% of the analysed samples. The reason for this high discrepancy could be due to the weak inhibitory effects of Pb²⁺ on urease activity which has been attributed to its poor affinity for the enzyme. This may have resulted in less sensitivity of the inhibition system towards Pb²⁺.

For the analysis of the multi-metal synthetic samples, AAS was first used to estimate the concentrations of the individual metals then the sum of the measured concentrations was taken as the total concentration (Table 10). The inhibition assays were used to estimate the concentration of the samples (as total heavy metals present) based on the respective inhibition values. For the inhibition assays, all the samples were diluted to a factor 10⁻¹ before measurable responses could be obtained. Comparisons of the total concentrations estimated via the two methods revealed about 70% significant differences. It appears the inhibition assay does not give full quantitative information with respect to multiple metals in solution; it seems to give a sum value of the heavy metal related toxicity of the samples. The results obtained via the assays were all lower than those obtained through AAS analysis thus giving further credence to our earlier observation that the inhibitory effect of mixed metals on urease activity is not additively synergetic.

Table 11: Comparative evaluation of two methods (urease Inhibition Assay and Atomic Absorption Spectroscopy) used to determine Heavy Metal Contents of synthetic water Samples

Heavy metals	Samples	Atomic absorption spectroscopy concentration (mg/L)	Urease inhibition assay concentration (mg/L)	T-test values
Copper	Cu-A	0.295 ± 0.05	0.278 ± 0.04	0.460
	Cu-B	0.694 ± 0.12	0.667 ± 0.02	0.384
	Cu-C	1.119 ± 0.02	1.114 ± 0.13	0.066
	Cu-D	1.063 ± 0.03	1.096 ± 0.09	0.603
	Cu-E	2.119 ± 0.04	1.989 ± 0.07	2.792*
	Cu-F	2.271 ± 0.01	2.183 ± 0.10	1.517
	Cu-G	0.994 ± 0.03	0.975 ± 0.03	0.776
	Cu-H	2.244 ± 0.15	2.238 ± 0.06	0.064
	Cu-I	0.988 ± 0.06	0.984 ± 0.03	0.103
	Cu-J	0.006 ± 0.00	ND	NA
Arsenic	As-A	1.734 ± 0.04	1.711 ± 0.12	0.315
	As-B	1.558 ± 0.02	1.596 ± 0.04	1.472
	As-C	1.645 ± 0.01	1.622 ± 0.05	0.781
	As-D	2.292 ± 0.06	2.268 ± 0.11	0.332
	As-E	2.223 ± 0.16	2.205 ± 0.14	0.146
	As-F	2.255 ± 0.08	2.244 ± 0.06	0.191
	As-G	2.268 ± 0.03	2.166 ± 0.03	4.164*
	As-H	1.897 ± 0.05	1.877 ± 0.02	0.643
	As-I	1.861 ± 0.02	1.850 ± 0.06	0.301
	As-J	1.862 ± 0.12	1.863 ± 0.09	0.012
Cadmium	Cd-A	1.156 ± 0.09	1.131 ± 0.03	0.456
	Cd-B	1.00 ± 0.01	0.975 ± 0.05	0.849
	Cd-C	2.582 ± 0.06	2.413 ± 0.05	3.748*
	Cd-D	1.850 ± 0.04	1.895 ± 0.10	0.189
	Cd-E	3.216 ± 0.03	3.231 ± 0.11	0.228
	Cd-F	2.242 ± 0.09	2.255 ± 0.06	0.208
	Cd-G	1.508 ± 0.07	1.480 ± 0.02	0.666
	Cd-H	0.230 ± 0.03	0.222 ± 0.02	0.348
	Cd-I	0.247 ± 0.01	0.244 ± 0.03	0.164
	Cd-J	0.140 ± 0.00	0.131 ± 0.07	0.221
Chromium	Cr-A	1.770 ± 0.03	1.735 ± 0.04	1.212
	Cr-B	0.198 ± 0.02	0.180 ± 0.09	0.338
	Cr-C	0.249 ± 0.01	0.219 ± 0.06	0.854
	Cr-D	ND	ND	NA
	Cr-E	0.108 ± 0.07	0.106 ± 0.11	1.099
	Cr-F	ND	ND	NA
	Cr-G	1.837 ± 0.05	1.800 ± 0.03	1.099
	Cr-H	1.972 ± 0.04	1.775 ± 0.11	2.915*
	Cr-I	0.766 ± 0.06	0.621 ± 0.04	3.723*
	Cr-J	0.194 ± 0.04	0.125 ± 0.05	1.867
Zinc	Zn-A	1.788 ± 0.08	1.684 ± 0.11	1.32
	Zn-B	4.631 ± 0.04	4.352 ± 0.12	3.82*
	Zn-C	1.957 ± 0.02	1.897 ± 0.05	1.93
	Zn-D	5.273 ± 0.09	5.163 ± 0.15	1.09
	Zn-E	1.800 ± 0.05	1.773 ± 0.10	0.42
	Zn-F	11.625 ± 0.13	11.434 ± 0.22	1.29
	Zn-G	10.395 ± 0.10	10.312 ± 0.09	1.07
	Zn-H	1.898 ± 0.06	1.822 ± 0.08	1.32
	Zn-I	3.863 ± 0.03	3.678 ± 0.06	4.78*
	Zn-J	6.685 ± 0.07	6.498 ± 0.13	2.19

Table 11 Cont.d

Heavy metals	Samples	Atomic absorption spectroscopy concentration (mg/L)	Urease inhibition assay concentration (mg/L)	T-test values
Lead	Pb-A	17.43 ± 0.05	17.084 ± 0.11	4.96*
	Pb-B	2.68 ± 0.01	2.445 ± 0.17	2.39*
	Pb-C	1.107 ± 0.02	1.007 ± 0.03	4.80*
	Pb-D	ND	ND	NA
	Pb-E	1.56 ± 0.04	1.513 ± 0.05	1.271
	Pb-F	ND	ND	NA
	Pb-G	1.39 ± 0.06	1.362 ± 0.15	0.3
	Pb-H	0.007 ± 0.00	ND	NA
	Pb-I	ND	ND	NA
	Pb-J	2.73 ± 0.01	2.617 ± 0.03	4.61*
Mixed-metals	MM-A	31.185±1.65	25.897 ± 1.35	3.48*
	MM-B	20.876±1.16	18.662 ± 0.83	2.69
	MM-C	42.878±2.74	40.473 ± 1.94	1.24
	MM-D	32.058±1.47	27.084 ± 1.16	4.60*
	MM-E	45.132±1.37	40.257 ± 2.15	3.31*
	MM-F	35.353±0.79	31.727 ± 0.93	5.15*
	MM-G	35.546±0.52	28.902 ± 0.85	11.6*
	MM-H	47.827±1.56	44.468 ± 2.60	1.92
	MM-I	41.618±1.02	37.558 ± 1.21	4.44*
	MM-J	37.533±0.55	33.437 ± 1.30	5.03*

IV. Conclusion

Inhibition studies are the first step towards developing an enzyme-inhibition based system for detection of heavy metals. Here we have carried out inhibition assays for six heavy metals using urease extracted from *Glycine max* seeds. All the investigated heavy metals exhibited linearly related dose-dependent inhibition of urease activity within varying concentrations. This factor demonstrates the feasibility of using the extracted urease for quantitative determination the metals. Application of the inhibition assay to single-metal synthetic samples gave satisfactory results when compared with atomic absorption spectroscopy (AAS) analysis. However, for the analysis of multi-metal samples, wide discrepancies were observed. This can be attributed to the pattern of inhibition by the mixed – metals which showed non-additive behaviour. Being inhibition based, the method under consideration cannot differentiate between inhibitors of the same class therefore it cannot discern which metals are present in a particular sample; it can only give a sum parameter of the overall heavy metal-associated toxicity of a sample. It would be useful as a screening tool to assess heavy metal contamination in water samples.

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