

Levels of Heavy Metals in Ground Water from Chantan Dump Site in Ghana

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Abstract: A study was undertaken to determine the concentrations of heavy metals in the ground water from landfill leachiest in Chantan, a suburb of Accra. Six wells were excavated from the Dump site and the concentrations for heavy metal such as copper, chromium, arsenic and zinc were monitored using wet digestion technique. HCl and HNO₃ were used followed by atomic absorption spectrophotometry. The heavy metals were quantified using external calibration with standards. Copper had the highest mean concentration of 1.410 mg/l in water from well three (3). This is above the maximum admissible and desirable limit set by the World Health Organization (WHO), an international organization. Chromium recorded the minimum mean concentration of 0.002 mg/l and this was found in wells (2) and (3). Generally, the study revealed that 83.3 % of water samples had mean concentrations below the WHO maximum admissible and desirable limit.

Keywords: Pollution: Heavy metal: Atomic Absorption spectrophotometry: Ground water: World Health Organization.

I. Introduction

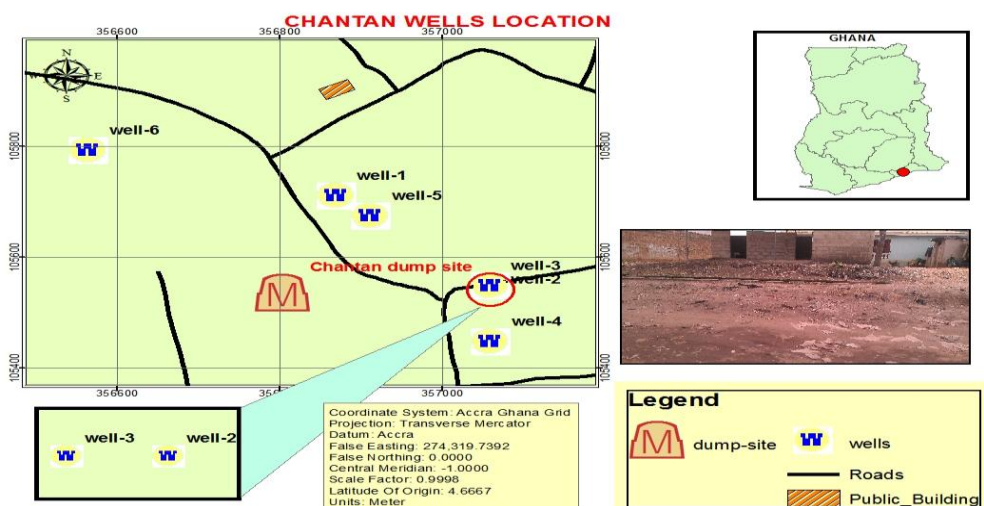
Groundwater accounts for about 98% of the world's fresh water resources and it is well distributed throughout the earth (Bouwer, 2002). Over the years, groundwater has served as a potential source of water supply. According to Ajibade et al., (2011), 90% of the population depend largely on hand-dug wells. However, the quality of ground water has been affected by human activities. More than one billion people in the world do not have suitable drinking water and statistics show that about three to five millions die annually from water-related diseases (Kumar and Singh, 2006). In the hydrological cycle, less than 0.1% of the metals are actually dissolved in the water and more than 99.9% are stored in sediments and soils (Karbassi et al., 2007, Pradit et al., 2010). The migration of heavy metals in groundwater is largely controlled by pH, amount of metal, cation exchange (Martinez and Motto, 2000), organic carbon content, oxidation state of mineral components as well as the redox potential of the ground water system (Sherene, 2010).

Heavy metals are among the major pollutants of water sources (Marcovecchio et al., 2007). The contamination of ground water by heavy metals has also assumed great significance during recent years due to their toxicity, bio concentration and bio accumulation in food chain. Heavy metals possess serious effects on human health and might cause various symptoms depending on the type and amount of the metal involved (Adepoju-Bello and Alabi, 2005). Due to the toxicity of heavy metals in water, guidelines for heavy metals in drinking water have been set by different International organizations such as WHO, EPA and the European Union Commission. Monitoring programs have therefore been established to regulate heavy metal concentrations in ground water sources, ensure compliance with national and international law, and to reassure consumers that drinking water is safe for consumption. There is need to continuously assess the quality of ground water sources to ensure that their concentrations conform with international standards (Ehi-Eromosele and Okiei, 2012).

II. Methodology

2.1 The study area

Chantan is situated in Greater Accra region of Ghana. Its geographical coordinates are 5° 37' 0'' North, 0° 15' 0'' West. Chantan market which used to be an open dumpsite was shut down in 2009 due to urbanization. The geology of the study area is made of Acrisols which have Ph less than 7.



2.2 Reagents and apparatus

Hydrochloric acid and nitric acid were analytical grade from BDH England, hydrogen peroxide was analytical grade from MERCK Darmstadt, Germany. deionised water, certified calibration standards (BDH England), 250 ml plastic bottles, flame atomic absorption spectrophotometer (Perkin Elmer 5100 PC), Whatman 4 filter paper (0.45 μm), Ph meter, conductivity meter and temperature probe (vernier software and technology).

2.3 Determination of p H, electrical conductivity and temperature

The electrical conductivity electrode was calibrated using sodium chloride solution of 12880 $\mu\text{S}/\text{cm}$. It was rinsed with deionised water and immersed into 50 ml of the water sample. The electrode was slowly moved in a circle for one minute until the digital readout stabilized. The probe was placed in the samples and the meter set to measure temperature. The process was repeated three times for all other water samples. For Ph measurements the electrode was calibrated with sodium chloride buffer solution at ph of 4.0, 7.0 and 9.0 respectively. The electrode was rinsed with deionised water and then immersed into 50 ml water samples. Measurements were allowed to stabilize and recorded.

2.4 Sampling

The study was conducted between February and March 2015. Water samples were taken from six wells within the sampling area beginning from the least suspected contaminated area. Prior to sample collection, all bottles were washed with detergents, dilute nitric acid and distilled water and dried in an oven. Water samples were collected using a rubber bucket tied to a rope. About 5 L of bulk sample was collected from all the six wells in a large plastic bowl after agitation of water to obtain a homogenous sample. Before taking the water samples, the 250 ml plastic bottles were rinsed three times with the water to be collected. Water samples were transferred into the sterilized 250 ml plastic bottles. About 2 ml of 10% HNO_3 was added to each bottle as a preservative (Singh *et al.*, 2010). The water samples were labeled and kept in an ice chest and then transported to the laboratory for storage in a refrigerator at about 4 $^{\circ}\text{C}$ before analysis.

2.5 Laboratory analysis

The water samples were digested using concentrated nitric acid according to Zhang (2007). 5 ml of the digested sample was filtered through a Whatman filter paper (0.45 μm) and diluted to 100 ml with deionised water and stored at 4 $^{\circ}\text{C}$ for analyses.

2.6 Analysis of anions

Chloride ions concentration in the water samples was ascertained by titration using 0.27 M K_2CrO_4 indicator and 0.1 M AgNO_3 . Sulphate ion solution was converted to barium sulphate and analysed by spectrophotometer at wavelength of maximum absorbance of 420 nm (USEPA, 1983). The concentration of the sulphate ions was then determined from a calibrated curve prepared from standard Na_2SO_4 solution. Nitrate ion concentration in the water samples were determined with the aid of spectrophotometer using the Cadmium reduction method (HACH, 1992).

2.7 Preparation of standards

Mixture of standard solution of selected metals (MIX 1) was prepared for Zn, As, Cu and Cr of concentration 1000 mg/l by dissolving their high purity metal salts in 0.5% HNO₃. Serial dilutions were made from stock solution to produce calibration standards of concentrations 0.2, 0.4, 0.6 and 0.8 mg/l respectively which were labeled as standard (MIX 1, MIX2, MIX 3 and MIX 4)

Table 1: Results of concentrations of heavy metals calibration standards

STANDARD	CONCENTRATION mg/l	ABSORBANCE
MIX 1	0.2	0.025
MIX 2	0.4	0.049
MIX 3	0.6	0.073
MIX 4	0.8	0.995

2.8 Instrumental analysis

The concentrations of Zn, As, Cu and Cr were determined with atomic absorption spectrophotometer (Perkin Elmer 5100 PC) using standard guidelines by APHA, AWWA, WEF 2001. A calibration curve was constructed and the concentration equivalent to the absorbance of the sample was read from the curve and recorded accordingly.

2.9 Quality assurance

Blanks were prepared and carried throughout the entire sample preparation and analytical process to determine whether samples were contaminated. Replicate samples were used to determine precision of results. Samples were spiked with 0.2, 0.5 and 1.0 mg/l of standard MIX solution to determine the recovery.

$$\text{Recovery} = \frac{\text{Amount of analyte recovered}}{\text{Amount of analyte spiked}} \times 100$$

III. Results and Discussion

The fate and transport of a metal in groundwater depends significantly on the chemical form and speciation of the metal (Allen et al., 1991). The mobility of metals in ground-water systems is hindered by reactions that cause metals to be adsorbed. These mechanisms can retard the movement of metals and also provide a long-term source of metal contaminants (NRC, 1994). While the various metals undergo similar reactions in a number of aspects, the extent and nature of these reactions varies under particular conditions. After discharge to an aquatic environment metals are partitioned between solid and liquid phases. Partitioning is affected strongly by variations in pH, redox state, organic content, and other environmental factors (Salomons, 1995). Chromium concentrations were above WHO limits of 0.005 mg/l in well 1 at ph 6.1 (Table 2 and 3). Most heavy metals are adsorbed at high Ph. However, literature reveals that ph range of 4 to 5 decreases the solubility and migration of chromium in water (Chrotowski et al., 1991) which may be due to presence of clay, iron and aluminium oxides as well as organic matter content. Most of chromium released into natural waters is particle associated, however, and is ultimately deposited into the sediment (Smith et al, 1995). In contaminated areas, Cr concentrations are commonly 300 to 500 mg/l (CRWQCB, 2000; Maxwell, 1997) and have been reported to reach 14 g/l (Palmer and Wittbrodt, 1991)

Table 2: Comparison of physical parameters of groundwater samples with WHO standards

LOCATION	MEAN pH	MEAN ELECTRICAL CONDUCTIVITY $\mu\text{S/cm}$	MEAN TEMPERATURE $^{\circ}\text{C}$
WELL 1	6.1	1200	28.5
WELL 2	6.8	1250	33
WELL3	6.5	1000	31
WELL 4	6.7	1500	32
WELL5	6.9	1620	31.8
WELL 6	6.9	1400	30.5
WHO LIMITS(2011)	6.5-8.5	500	

Table 3: Results of heavy metal concentrations in ground water samples from six wells in Chantan (mg/l)

WELL	MEAN CONCENTRATION \pm SD in mg/L			
	Cr	As	Zn	Cu
1	0.081 \pm 0.006	0.108 \pm 0.008	0.008 \pm 0.001	0.903 \pm 0.042
2	0.002 \pm 0.001	0.012 \pm 0.010	0.009 \pm 0.001	1.073 \pm 0.042
3	0.002 \pm 0.001	0.118 \pm 0.026	0.006 \pm 0.001	1.410 \pm 0.036
4	0.041 \pm 0.010	0.007 \pm 0.008	0.006 \pm 0.001	0.873 \pm 0.025
5	0.009 \pm 0.001	0.007 \pm 0.002	0.006 \pm 0.002	0.850 \pm 0.075
6	0.005 \pm 0.001	0.008 \pm 0.001	0.009 \pm 0.001	0.938 \pm 0.035
WHO LIMIT (2011)	0.050	0.010	5.000	2.000

Concentrations of As in well 1, 2 and 3 were relatively higher than recommended values of 0.01 mg/l by WHO (Table 3). This may be attributed to increase in temperature of water which facilitates the migration of heavy metals in water. When the temperature increases metal activity in the water may be high and has faster migration rates. Yu Ling et al., (1998) concluded that considerable enrichment of heavy metals in the sub-micron range occurs during vaporization of heavy metals in the high temperature environment and the subsequent condensation in lower temperature regions of the systems. In addition, arsenic mobility increases as pH increases (Smith et al., 1995). The effective desorption of As from adsorbent could be achieved using a mixed solution of Cl^{-1} and OH^{-1} (Elizalde-Gonzalez et al., 2001; Lenoble et al., 2002). Arsenic can readily combine with other elements, and inorganic arsenic compounds are formed when arsenic combines with elements such as chloride and sulfur compounds (OSHA, 2004). The elevated concentrations of As were related to high Cl^{-1} and SO_4^{-2} in water (table 4). Erickson and Barnes (2005) stated that in the upper Midwest, USA, elevated arsenic concentrations in public drinking water systems, where twelve percent of public water systems were located exceed 10 $\mu\text{g/l}$, the USEPA drinking water guideline value for arsenic. Recently, it has been reported that in Mexico, natural groundwater As contamination ranked 0.5 to 3.7 mg/ l (Hossain, 2006). Marshall et al., (2007) found that drinking water in region II of Chile is supplied mainly by rivers that contain inorganic arsenic at very high concentrations. Kelepertsis et al., (2006) found elevated concentrations of As (125 $\mu\text{g/l}$) in the drinking water of Eastern Thessaly, Greece, while recently, Jovanovic et al. (2011) found that 63 % of all water samples exceeded Serbian and European standards for arsenic in drinking water.

Table 4: Mean concentration of anions in six wells at Chantan

MEAN CONCENTRATION ± SD OF ANIONS			
WELL	Cl^{-1}	SO_4^{-2}	NO_3^{-1}
1	352.37±10.21	490.28±5.80	1.98±0.05
2	308.53± 7.80	267.67± 9.07	1.65±0.31
3	411.13±9.92	607.50±6.14	1.57±0.03
4	264.30±8.22	324.33±6.03	1.70±0.03
5	316.97±6.21	439.33±5.13	1.59±0.02
6	517.23±7.51	264.67±5.51	1.44±0.02

The low concentrations of As in well 4, 5 and 6 may be due adsorption and co precipitation with hydrous iron oxides which are the removal mechanisms under most environmental conditions (Krause and Ettel, 1989; Pierce and Moore, 1982).

Concentrations of Cu and Zn were below the WHO limits of 5.0 and 2.0 mg/l respectively for all the wells samples in Chantan. Howell and Gawthorne (1987) discussed the important role of organic matter in the retention and behavior of copper. Igloria et al. (1997) included the possibility of formation of complexes between heavy metals and organic matter colloids in solutions, as well as those attached to soil surfaces. Sposito (1986), on the other hand, claimed that the effect of these colloids might retard metals transport. Clay carbonates, or manganese and hydrous oxides, readily adsorb zinc (Smith et al., 1995). The greatest percentage of total zinc in polluted ground water and sediment is associated with iron (Fe) and manganese (Mn) oxides. As with all cationic metals, zinc adsorption increases with pH and this is evident in table 1. Xu et al. (2006) reported that the mean concentrations of Cu and Zn in drinking water from Shanghai, China were 0.29 mg/l and 0.91 $\mu\text{g/l}$, respectively; which were lower than USEPA and WHO guideline values. On the contrary, high concentrations of Cu (88 to 147 $\mu\text{g/l}$) were found in bottled drinking waters sold in Canada (Dabeka et al., 2002).

IV. Conclusion

From this study it can be concluded that 83.3 % of heavy metals analyzed were below the WHO permissible levels for drinking water. Copper had the highest mean concentration of 1.410 mg/l in water from well three (3). Chromium recorded the minimum mean concentration of 0.002 mg/l and this was found in wells (2) and (3). There was a positive correlation for Ph, temperature and conductivity with the concentrations of heavy metals recorded in wells in Chantan Dump site.

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References

- [1]. Adepoju-Bello, A.A. and O.M. Alabi, 2005. Heavy metals: A review. The Nig. J. Pharm., 37: 41-45.
- [2]. Ajibade, O. M., Omisanya, K. O. And Odunsi, G. O., (2011): Groundwater potability and flow direction of urban aquifer, Ibadan, Southwestern Nigeria; Journal of the Nigerian Association of Hydrogeologists, Volume 21, ISSN 0795-6495. Pp38-55 APHA-AWWA-WEF. (2001). Standard Methods for Examinations of Water and Wastewater, (22nd ed., pp. 4-146). Washington D.C.

- [3]. Allen, J.P and Torres, I.G. (1991), "Physical Separation Techniques for Contaminated Sediment," in Recent Developments in Separation Science, N.N. Li, Ed., CRC Press, West Palm Beach, FL, Vol V.
- [4]. Bower, H.(2002): Artificial recharge of groundwater; Hydrology and Engineering; *Hydrogeologist Journal* 10(1).pp121-142
- [5]. California Regional Water Quality Control Board (CRWQCB), 2000, Waste Discharge Requirements for In-Situ Pilot-Study for the Chemical Reduction of Cr. Order No. R1-2000-54.
- [6]. Chrotowski, P.C., Durda, J.L., and Gdelmann, K.J (1991). The use of natural processes for the control of chromium migration. *Remediation* vol1 no 3 pp341-351
- [7]. Dabeka, R., Conacher, H.B.S., Lawrence, J.F and Pepper, K. (2002). Survey of Bottled Drinking Waters Sold in Canada for Chlorate, Bromide, Bromate, Lead, Cadmium and Other Trace Elements. *Food Additives and Contaminants* 19(8):721-32 .
- [8]. Ehi-Eromosele C. and Okiei W. (2012). *Resources and Environment*,2(3):82-86
- [9]. Elizalde-Gonzalez, M.P., Mattusch, J., Einicke, W.D., Wennrich, R., 2001. Sorption on natural solids for arsenic removal. *Chem. Eng. J.* 81, 187-195.
- [10]. Erickson, M.L. and Barnes, R.J. (2005). Glacial Sediment Causing Regional-Scale Elevated Arsenic in Drinking Water. *Ground water* pp 1-10
- [11]. HACH, 1992. *Water Analysis Handbook*. Hach Company, Loveland, Colorado, USA, pp: 401-402, 427-428, 537-538.
- [12]. Hossain MF (2006). Arsenic contamination in Bangladesh - An overview. *Agro. Ecosystem Environ.* 113:1-16.
- [13]. Igloria, R.V., W.E. Hathhorn, Member, ASCE and D.R. Yonge. 1997. NOM and Trace Metal Attenuation During Storm-Water Infiltration. *Journal of Hydrologic Engineering*. Vol. 2, No. 3, pp. 120-127
- [14]. Jovanovic D, Jakovijevic B, Ragic-Milutinovic Z, Paunovic K, Pekovic G, Knezevic T (2011). Arsenic occurrence in drinking water supply systems in ten municipalities in Vojvodine Region, Serbia. *Environ. Res.* 111:315-318.
- [15]. Karbassi A., Ayaz G., Nouri J. 2007. Flocculation of trace metals during mixing of Talar river water with Caspian Seawater. *International Journal of Environmental Research*, 1(1), 66-73
- [16]. Kelepertsis A, Alexakis D, Skordas K (2006). Arsenic, antimony and other toxic elements in the drinking water of Eastern Thessaly in Greece and its possible effects on human health. *Environ. Geol.* 50:76-84.
- [17]. Krause, E. and Ettl V.A. (1989), "Solubilities and Stabilities of Ferric Arsenate Compounds," *Hydrometallurgy*, 22, pp. 311-337.
- [18]. Kumar R., Singh R.N. 2006. *Municipal water and wastewater treatment*. New Delhi: Capital Publishing Company
- [19]. Lenoble, V., Bouras, O., Deluchat, V., Serpaud, B., Bollinger, J.C., 2002. Arsenic adsorption onto pillared clays and iron oxides. *J. Colloid Interface Sci.* 255, 52-58.
- [20]. Marcovecchio, J.E., S.E. Botte and R.H. Freije, 2007. Heavy Metals, Major Metals, Trace Elements. In *Handbook of Water Analysis*. L.M. Nollet, 2 Ed. London: CRC Press, pp: 275-311
- [21]. Marshall G, Ferreccio C, Yuan Y, Bates MN, Steinmaus C, Selvin S, Liaw J, Smith H (2007). Fifty-year study of lung and bladder cancer mortality in Chile related to arsenic in drinking water. *J. Natl. Cancer I.* 99:920-928
- [22]. Martinez, C. E. and Motto, H. L. (2000). Solubility of lead, zinc and copper added to mineral soils, *Environmental Pollution*, 107, no. 1, pp. 153-158
- [23]. Maxwell, C.R. (1997), Investigation and remediation of Cr and nitrate groundwater contamination: case study for an industrial facility, *Journal of Soil Contamination*, 6, 6, 733-749.
- [24]. NRC (1994), *Alternatives for Ground Water Cleanup*, National Research Council, National Academy Press, Washington, D.C.
- [25]. Occupational Safety & Health Administration (OSHA). Arsenic. Retrieved 9/14/2011 from <http://www.osha.gov/SLTC/arsenic/index.html>
- [26]. Palmer, C.D. and Wittbrodt, P.R., 1991, Processes affecting the remediation of Cr contaminated sites, *Environmental Health Perspectives*, 92, 25-40.
- [27]. Pierce, M.L. and Moore, C.B. (1982), "Adsorption of Arsenite and Arsenate on Amorphous Iron Hydroxide," *Water Res.*, 16:1247-1253
- [28]. Pradit S., Wattayakorn G., Angsupanich S., Baeyens W., Leermakers M. 2010. Distribution of trace elements in sediments and biota of Songkhla Lake, Southern Thailand. *Water, Air, and Soil Pollution* 6(1), 155-174.
- [29]. Salomons W 1995. Environmental impact of metals derived from mining activities: Processes, predictions, prevention. *Journal of Geochemical Exploration* 52:5-23.
- [30]. Singh A., Sharma R. K., Agrawal M. and Marshall F. M. (2010). Risk assessment of heavy Metal toxicity through contaminated vegetables from wastewater irrigated areas in Varanasi, India. *Tropical Ecological Issue* 51. 2: 375-387.
- [31]. Sherene. T (2010). Mobility and transport of heavy metals in polluted soil environment. *Biological Forum — An International Journal*, 2(2): 112-121
- [32]. Smith, L.A., Means, J.L., Chen, A., Alleman, B., Chapman, C.C., Tixier, J.S., Jr., Brauning, S.E., Gavaskar, A.R., and Royer, M.D. (1995), *Remedial Options for Metals-Contaminated Sites*, Lewis Publishers, Boca Raton, FL.
- [33]. Sposito, G. (1989), *The Chemistry of Soils*, Oxford University Press, New York
- [34]. United States Environmental Protection Agency (USEPA), (1983). *Method of Chemical analysis of Water and Wastewater*, EPA/600/4-79/020, 374.3.1-375.4.3
- [35]. Xu P, Huang SB, Wang ZJ, Lagos G (2006). Daily intakes of copper, zinc and arsenic in drinking water by population of Shanghai, China. *Sci. Total Environ.* 362:50-55
- [36]. Yu-Ling-Wei and Yu Ling Huang. (1998). Behaviour of sequential extraction of lead from thermally treated lead (II) *Doped Alumina J. Environ. Qual.* 27: 343-348
- [37]. Zhang., C. (2007). *Fundamental of environmental sampling and analysis*. Wiley, New York, p 109